

1 **Wheat is an emerging exposure route for arsenic in Bihar, India**

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18 **Abstract:**

19 In arsenic (As) endemic areas of south-east Asia, where a subsistence rice-based diet is
20 prevalent, As exposure from food is mainly focused on rice intake. However, consumption of
21 wheat is substantial and increasing. We present a probabilistic assessment of increased cancer
22 risk from wheat-based food intake in a study population of rural Bihar, India where As exposure
23 is endemic. Total As in wheat grains ($43.64 \pm 48.19 \mu\text{g}/\text{kg}$, $n=72$) collected from 77 households
24 across 19 villages was found to be lower than reported As in wheat grains from other south-
25 east Asian countries but higher than a previous study from Bihar. This is the first study where

26 As concentration in wheat flour was used for risk estimation, bearing in mind that it was the
27 flour obtained after indigenous household processing of the grains that was used for making
28 the home-made bread (*chapati*) which contributed 95% of wheat intake for the studied
29 population. Interestingly, while 78% of the surveyed participants (n=154) consumed rice every
30 day, *chapati* was consumed every day by 99.5% of the participants. In contrast to previous
31 studies, where As concentration in wheat grain was found to be lower than the flour due to the
32 removal of the bran on grinding, we did not find any appreciable lowering of arsenic in the
33 wheat flour ($49.80 \pm 74.08 \mu\text{g/kg}$, n=58), most likely due to external contamination during
34 processing and grinding. Estimated gender adjusted excess lifetime cancer risk of 1.23×10^{-4} for
35 the studied rural population of Bihar indicated risk higher than the 10^{-4} - 10^{-6} range, typically
36 used by the USEPA as a threshold to guide regulatory values. Hence, our findings suggest As
37 exposure from wheat-based food intake to be of concern not only in As endemic areas of rural
38 Bihar but also in non-endemic areas with similar wheat-based diet due to public distribution of
39 the wheat across India.

40

41 **Key words:** arsenic; wheat flour; probabilistic risk; wheat intake; Bihar

42 **1. Introduction:**

43 Arsenic (As) contamination in groundwater of Bihar, India was first reported in 2003
44 (Chakraborti et al., 2003), with more than 9 million people currently facing health risks due to
45 As exposure (Chakraborti et al. 2017). In a 2007 study, 10.8% (7,164 out of 66,623) of drinking
46 water samples, covering 11 districts of Bihar, were reported to have As concentration higher
47 than the Indian permissible limit of $50 \mu\text{g/l}$ (Nickson et al., 2007). But in 2017, Chakraborti et
48 al., reported that 17.8% of the drinking water samples (out of 19,961) had As above $50 \mu\text{g/l}$.

49 Moreover, in a recent study 22 out of 38 districts of Bihar were reported to have As in drinking
50 water above the WHO permissible limit of 10 $\mu\text{g/l}$ (Chakraborti et al., 2018).

51 Like all other As affected areas in south-east Asia, exposure is no longer restricted to drinking
52 water and food is becoming a significant route. While exposure from rice is well explored
53 (Mondal and Polya 2008; Mondal et al., 2010; Mwale et al., 2018) and health effects (Banerjee
54 et al., 2013) along with perception of risk (Mondal et al, 2019) studied, little is known about
55 exposure from wheat-based food intakes especially in India and only few studies have reported
56 As concentrations in Indian wheat samples. This is mostly because, unlike rice which is
57 cultivated in submerged soil conditions resulting in high As uptake, wheat is not reported to
58 have As as high as in rice (Bhattacharya et al., 2010). Kumar et al. (2016) reported mean As
59 concentration of $27 \pm 24 \mu\text{g/kg}$ ($n=35$, range= $7.7-108 \mu\text{g/kg}$) in wheat grain samples collected
60 from households in the Samastipur district of Bihar, India, while Bhattacharya et al. (2010)
61 previously reported much higher mean As concentration of $129 \mu\text{g/kg}$ ($n=8$, range= $90-200$
62 $\mu\text{g/kg}$) in wheat samples collected from fields of West Bengal, India. Arsenic was also reported
63 in wheat samples collected from fields of Ropar wetland, Punjab (mean= $110 \pm 20 \mu\text{g/kg}$;
64 range= $30-210 \mu\text{g/kg}$; $n=9$), an area not known to have geogenic As exposure, and authors
65 largely attributed the As contamination to addition of industrial wastes, excessive use of
66 pesticides and fertilizers in agricultural fields, and settlement of fly ash on soil (Sharma et al.,
67 2016).

68 With greater annual consumption than rice, wheat is the important staple food worldwide
69 (Rasheed et al., 2017). It is also the most important food-grain of India, next to rice but the
70 exposure of As from wheat is sparsely explored. Along with variation in As concentrations in
71 samples collected from different regions of India there exists variation in the pattern of dietary
72 consumption among different sub-populations of the different As exposed areas of the country.
73 These variations were rarely considered for As health risk assessment.

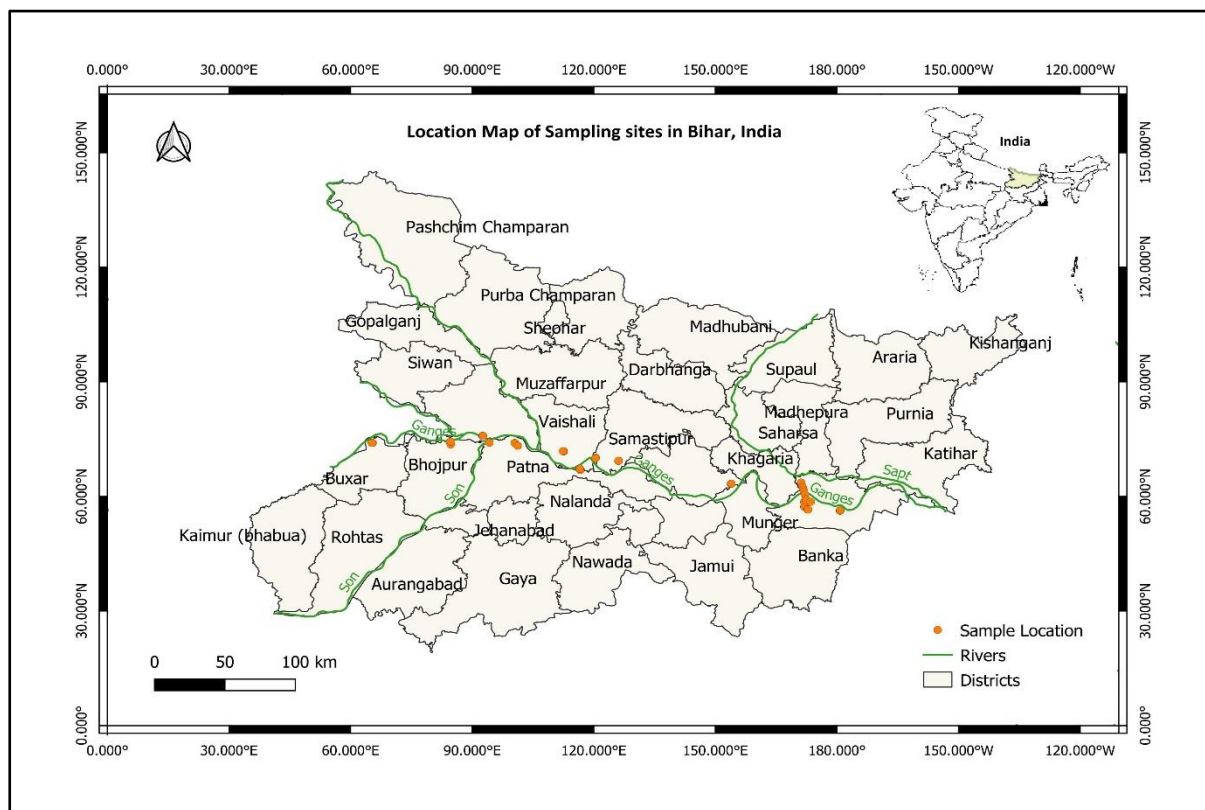
74 In this study, whole wheat grain and wheat flour samples were collected from households of
75 As affected areas in Bihar and the total As was determined in both samples. We focused on
76 wheat-based food intake and determined the overall increased cancer risk due to As in wheat
77 using a probabilistic method. To best of our knowledge, this is the first study where detailed
78 dietary assessment was done to estimate the As exposure from wheat-based food intake and
79 furthermore As in wheat flour was considered over wheat grain for risk assessment.

80 **2. Materials and methods**

81 *2.1 Survey area*

82 As a part of an on-going study “Nature and nurture in arsenic induced toxicity of Bihar, India”,
83 wheat samples were collected from 77 households across 19 villages in 10 out of 22 As affected
84 districts of Bihar (Figure 1) in 2017-18. Study locations were selected based on the previous
85 data on arsenic affected areas in Bihar, India. In each village, the survey team met with either
86 the village head or local elderly available resident to gather information on the village
87 demography and then randomly selected households stratified by low and high economic status
88 based on initial observation including type of house. From each household, an adult male and
89 female took part in a questionnaire survey (n=154) administered by the research team which
90 included a detailed food frequency questionnaire (FFQ) combined with 24-hour recalls
91 (24HRs) (Freedman et al., 2018) to determine the total wheat consumption (g/day). The
92 percentage contribution of each wheat-based food consumed towards total wheat consumption
93 was determined using 24HRs and the information on composite mean nutrient in each food as
94 per the latest Indian food composition database (Longvah et al., 2017). This study which is a
95 part of the an on-going research project that includes further sample and data collection, for
96 example, self-recorded health status, lifestyle, demographic and socio-economic conditions
97 along with anthropometric measurements was conducted in accordance with national and

98 international guidelines for the protection of human subjects and was approved by both the
99 University of Salford Ethics Committee and Mahavir Cancer Sansthan Institutional Ethics
100 Committee.



101

102 **Figure 1:** Sampling location of households surveyed in Bihar, India

103 **2.2 Sample collection and analysis**

104 A total of 72 wheat grain and 58 wheat flour samples (53 paired samples) were collected from
105 the 77 households. Wheat grain and wheat flour samples were collected in plastic zip-lock bags
106 and transported to the laboratory in Patna, stored at room temperature until the grains were
107 washed, dried, and ground to powder. Powdered grain and flour samples were shipped to the
108 University of Newcastle, Australia for analysis.

109 Both powdered wheat grain and flour samples were digested for the analysis of total As and
110 other elements (cadmium (Cd), calcium (Ca), copper (Cu), lead (Pb), magnesium (Mg),

111 manganese (Mn), selenium (Se), and zinc (Zn)), based on the protocol of Roychowdhury et al.
112 (2002). The determination of As and other trace elements was carried out with an Agilent 7900
113 (Tokyo, Japan) inductively coupled plasma mass spectrometer (ICP-MS). Major elements such
114 as Ca, Mg, Zn were analysed using the dual view (Axial and radial) inductively coupled plasma
115 emission spectrometer (ICP-OES, PerkinElmer Avio 200). CRM, blanks, duplicates, and
116 continuing calibration verification (CCV) were included in each batch throughout the elemental
117 analysis.

118 **2.3 Risk assessment**

119 Overall increased cancer risk was calculated using the following equation (USEPA, 1989)

$$120 \quad TR = CSF \times \frac{(IR \times C \times C_{ing} \times BCF) \times ED}{BW \times LT}$$

121 Where TR is the excess lifetime cancer risk; CSF is the oral cancer slope factor for inorganic
122 As (per mg/kg/day); IR is the ingestion rate of the wheat based food intake (g/day) estimated
123 by combining information from 24HRs and FFQ for every participant (n=154); C is the total
124 As concentration in the wheat ($\mu\text{g}/\text{kg}$); C_{ing} is the proportion of inorganic As in the wheat;
125 BCF is the bioconcentration factor; ED is the exposure duration (years); BW is body weight of
126 the exposed person (kg); LT is the life expectancy of the exposed person.

127 We used both point estimates and a probabilistic approach for increased lifetime cancer risk
128 estimation. For this later risk assessment approach, a Monte Carlo simulation of 10,000
129 iterations was carried out. Of the input variables in the above-mentioned equation, C, IR and
130 BW were determined in this study, the remaining parameters were estimated from analogous
131 studies.

132 Data was analysed using Microsoft Excel and Stata 11.2 for descriptive statistics and the
133 probability distributions for input variables were determined with the software @Risk (Trial

134 Version 7, Palisade Corp., USA) in combination with Microsoft Excel for probabilistic risk
135 estimation.

136

137 3. Results and discussion

138 3.1 Arsenic concentration in wheat grain and flour samples

139 The percentage recovery for rice flour (SRM 1568a) and mean percentage difference between
140 the duplicates (n=12) are shown in Table 1. Mean total As concentration in wheat grain (43.64
141 ± 48.19 $\mu\text{g}/\text{kg}$; range 0.96 - 234.3 $\mu\text{g}/\text{kg}$; Table 2), was lower than concentrations reported in
142 previous studies especially from Bangladesh, China and Pakistan (Table 3, Figure 2) but higher
143 than the reported concentrations in wheat from a previous study in Bihar (27 ± 24 $\mu\text{g}/\text{kg}$; range
144 7.7 - 108 $\mu\text{g}/\text{kg}$) by Kumar et al., (2016). This could be due to the type of survey, for example,
145 household or market-based compared to field study; the wheat variety; and most importantly
146 the study area being highly As contaminated or not. In compiling the studies (Table 3 and
147 Figure 2) attention was paid to the quality assurance of As analysis in individual studies as
148 summarised in the supplementary material.

149 Wheat could be a decent source of several vitamins and minerals, including Se and Mn. Both
150 in wheat grains and flour, Pb, Se, Fe and Mg were found to be positively correlated with As,
151 while Mn was found to be negatively correlated (Table 2). Further studies on nutritional value
152 of wheat cultivated in As affected areas could help determine the effect of As on essential
153 nutrients in wheat.

154 **Table 1.** Concentrations of As and other elements in the CRM and percentage variation in
155 replicates analysed

156

| Element | NIST SRM 1568b (Rice flour) (n=5) | | | Variation in duplicates (n=12) |
|---------|--|---|-----------------|--------------------------------|
| | Certified value ($\mu\text{g}/\text{kg}$) | Observed value ($\mu\text{g}/\text{kg}$) | Recovery (%) | Mean \pm SD (%) |

| | | | | |
|----|--------------|-------------|-------|---------------|
| As | 285 ± 14 | 254 ± 15 | 89.2 | 14.36 ± 13.22 |
| Cd | 22.4 ± 1.3 | 17.9 ± 2 | 80.1 | 9.42 ± 9.25 |
| Pb | 8* ± 3 | 6.3 ± 0.8 | 78.7 | 29.04 ± 22.21 |
| Se | 365 ± 29 | 300 ± 27 | 82.4 | 38.69 ± 32.41 |
| Mn | 19200 ± 1800 | 18966 ± 665 | 98.8 | 4.73 ± 3.33 |
| Cu | 2350 ± 160 | 2890 ± 284 | 92.9 | 9.08 ± 9.37 |
| Zn | 19400 ± 260 | 14210 ± 293 | 73.2 | 4.15 ± 3.08 |
| Ca | 118.4 ± 3.1 | 126.2 ± 3.7 | 106.6 | 5.44 ± 3.07 |
| Fe | 7.42 ± 0.44 | 7.22 ± 0.76 | 97.2 | 11.46 ± 8.05 |
| Mg | 559 ± 10 | 470.3 ± 12 | 84.1 | 3.80 ± 3.32 |

157 *reference value only

158

159 **Table 2:** Total As and concentration of other elements in wheat grain and wheat flour samples
 160 collected from households in Bihar, India

| | Wheat grain (n=72) | | Wheat flour (n= 58) | |
|-------------------|----------------------|---------------------|---------------------|--------------------|
| | mean ± SD | Range | mean ± SD | Range |
| As (µg/kg) | 43.64 ± 48.19 | 0.96 - 234.3 | 49.80±74.08 | 3.59-448.25 |
| Cd (µg/kg) | 19.70 ± 9.78 | 4.35-51.41 | 21.57±8.60 | 5.03-51.61 |
| Pb (µg/kg) | 61.82 ± 59.54* | 10.31-384.73 | 62.82±32.66* | 19.35-178.53 |
| Se (µg/kg) | 138.41±168.53* | 0.67-953.67 | 113.52±111.91* | 0.67-432.66 |
| Mn (mg/kg) | 33.71± 6.7** | 16.73-58.51 | 32.84±7.39** | 13.90- 51.70 |
| Cu (mg/kg) | 4.54 ± 0.92 | 2.98-9.24 | 4.84±0.80 | 3.61-8.37 |
| Zn (mg/kg) | 23.44 ± 4.55 | 14.82-34.44 | 23.01±4.57 | 14.77-36.11 |
| Ca (mg/kg) | 425.40 ± 77.21 | 305.91-870.04 | 425.60±72.28 | 229.52-596.22 |
| Fe (mg/kg) | 83.65 ±113.52* | 25.10-779.83 | 67.70±33.32* | 33.20-180.34 |
| Mg (mg/kg) | 1085.98±102.39* | 896.22-1456.18 | 1100.37±131.01* | 555.03-1379.19 |

161 *elements that were positively correlated with As; **elements that were negatively correlated
 162 with As (using Spearman's rho (p <0.05))

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164 **Table 3:** Total As concentration in whole wheat grains in different studies

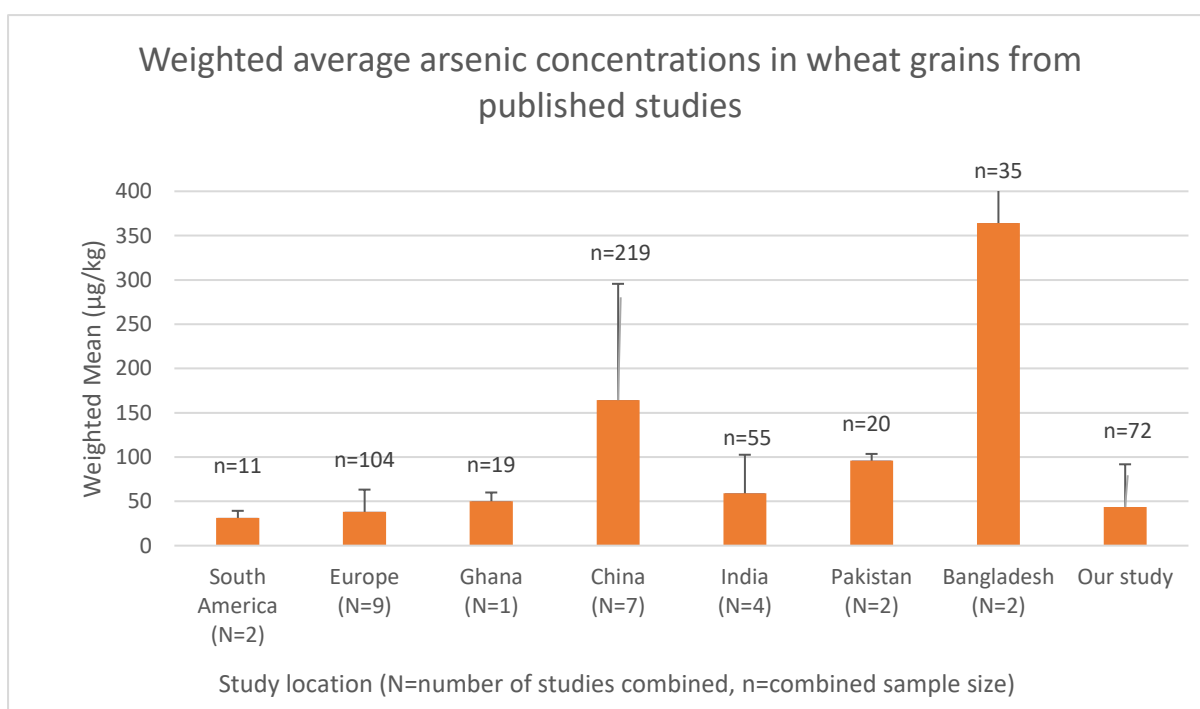
| Country | Area or site or type | Year (published) | Sample Size (n) | Mean total As (µg /kg) | (Range) SD (µg /kg) | Reference |
|---------|----------------------|------------------|-----------------|------------------------|---------------------|--------------|
| India | Gangetic, WB | 2005 | 3 | 94.52 | 22.79 | Samal (2005) |

| | | | | | | |
|------------|--------------------------------------|------|-----|------------|--------------------|----------------------------|
| Scotland | Field | 2007 | 29 | 30 | 10-210 | Williams et al. (2007) |
| China | Kunshan | 2008 | 40 | 38(median) | 29-86 | Huang et al. (2008) |
| England | Field | 2008 | 37 | 70 | 10-500 | Williams et al. (2007) |
| Pakistan | Manchar (water As > 10 ppb) | 2008 | 150 | 317 | 96 | Arain et al. (2009) |
| Pakistan | Manchar (water As <10 ppb) | 2008 | 150 | 175 | 56 | Arain et al. (2009) |
| China | Zhengzhou, Henan | 2009 | 40 | 136.66 | 70-210 | Liu et al. (2009) |
| India | Nadia, WB | 2010 | 8 | 129 | 10-190 | Bhattacharya et al. (2010) |
| Europe | Multinational | 2010 | 8 | 7.7 | 5.4 | Zhao et al. (2010) |
| Hungary | Martonvasar, 2006 | 2010 | 2 | 6 | 3 | Zhao et al. (2010) |
| Hungary | Martonvasar, 2007 | 2010 | 2 | 3 | 2 | Zhao et al. (2010) |
| Poland | Choryn, 2007 | 2010 | 2 | 13 | 3 | Zhao et al. (2010) |
| UK | Woolpit, 2007 | 2010 | 2 | 9 | 4 | Zhao et al. (2010) |
| Ghana | Market based (origin China) | 2011 | 19 | 50 | 10 | Adomako et al. (2011) |
| Italy | Field | 2010 | 141 | 8.62 | 4.6-14.8 | Cubadda et al. (2010) |
| Italy | Field | 2011 | 3 | 29.8 | 0.7 | D'Amato et al. (2011) |
| China | Huaibei, Coal Mining Area | 2013 | 75 | 33.3 | (6.5-54.9) 8.05 | Shi et al. (2013) |
| China | Mangolia (low As) | 2014 | 25 | 66.9 | 22.8-154 | Tong et al. (2014) |
| China | Mangolia (high As) | 2014 | 25 | 238 | 138-365 | Tong et al. (2014) |
| Bangladesh | All Agroclimatic Zone | 2015 | 30 | 281 | 1 | Ahmed et al. (2015) |
| Brazil | Sao Gotardo | 2015 | 3 | 19 | - | Corguinha et al. (2015) |
| Bangladesh | Faridpur | 2016 | 5 | 864 | 490-1150 | Kamrozzaman et al. (2016) |
| Pakistan | Field of 12 districts | 2016 | 12 | 90 | 63 | Al-Othman et al. (2016) |
| India | Field of Ropar wetland, Punjab | 2016 | 9 | 110 | 20 | Sharma et al. (2016) |
| India | Bihar | 2016 | 35 | 27 | (7.7 – 108) 24 | Kumar et al. (2016) |
| Argentina | Santa Fe | 2016 | 8 | 36 | <LOD - 73 | Sigrist et al. (2016) |

| | | | | | | |
|----------|-----------------------|------|----|-------|-----------------------|------------------------|
| Italy | North West Italy | 2018 | 18 | 10 | (06-22) | Giordano et al. (2018) |
| China | Henan | 2018 | 48 | 270 | 130-340 | Guo et al. (2018) |
| Pakistan | Four Region | 2018 | 8 | 105 | 61.47 | Rasheed et al. (2018) |
| China | Field Dongdagou | 2018 | 22 | 417 | 155 | Zhang et al. (2018) |
| China | Field Xidagou | 2018 | 14 | 224 | 279 | Zhang et al. (2018) |
| Belgium | Field | 2018 | 9 | 22 | 6 | Ruttens et al. (2018) |
| Bihar | Household based study | | 72 | 43.64 | (0.96-234.3) 48.19 | This study |

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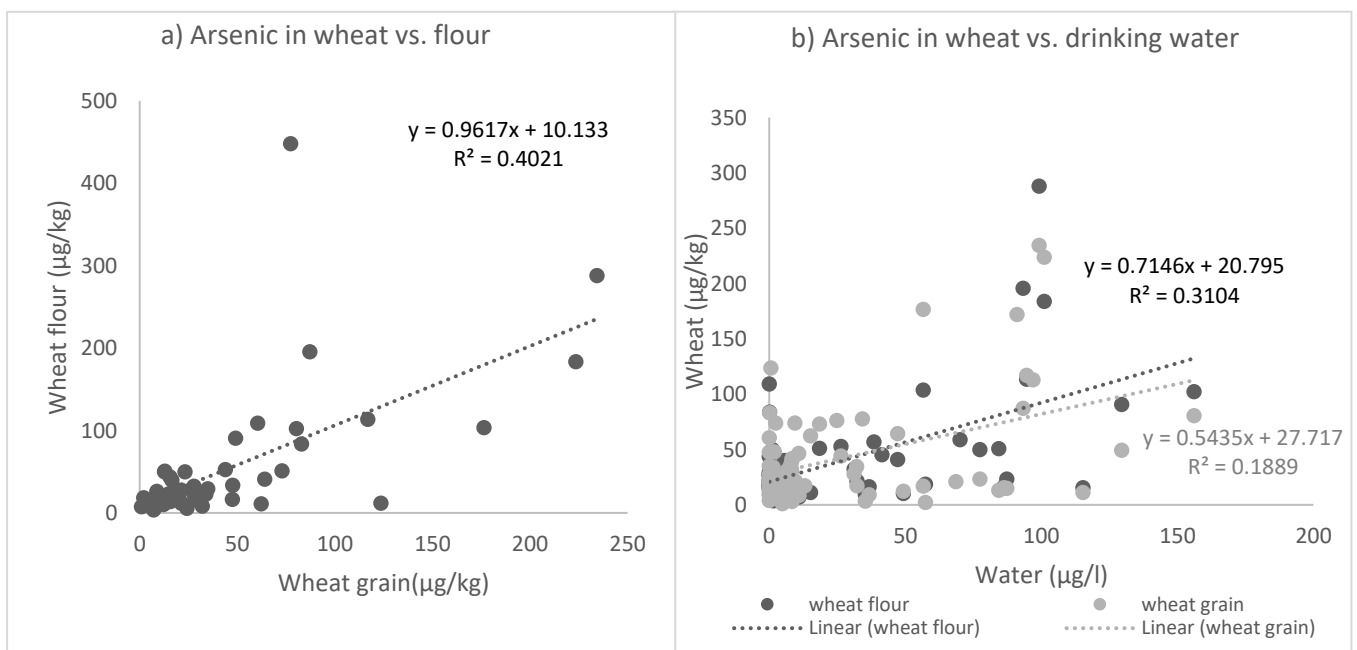
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168 **Figure 2:** Weighted mean and SD of As concentrations in wheat grains using 27 published
169 studies

170 Though the average As concentration in the wheat flour ($49.80 \pm 74.08 \mu\text{g/kg}$) was higher than
171 in wheat grains (Table 2) and correlated (Figure 3a) but based on an independent t-test run on
172 53 paired samples, there was no significant difference in As concentration between wheat grain
173 and wheat flour ($t(52) = 1.03, p = 0.31$) samples.

174 In previous studies, As concentration in flour was found to be lower. For example, by 40%
 175 (grain: 69.0 ± 17.0 ; flour: $41.2 \pm 13.7 \mu\text{g/kg}$) in the study by Zhao et al. (2010); by 56% (grain:
 176 29.8 ± 0.7 ; flour: $13.2 \pm 1.3 \mu\text{g/kg}$) in the study by D'Amato et al. (2011); and by 52% (grain:
 177 33.1 ± 7.9 ; flour: $15.8 \pm 3.7 \mu\text{g/kg}$) in the study by Zhang et al. (2009). This is because milling
 178 of wheat grain into bran and white flour fractions results in a three to four-fold higher As
 179 concentration in bran than in the white flour (Zhao et al., 2010), thus most of the whole grain
 180 As remains in the bran. But such significant decrease was not observed in this study and in
 181 fact an overall increase based on average concentrations in wheat grain and wheat flour was
 182 noted. This could be attributed to either soaking of the grains in As contaminated water (as
 183 shown in Figure 4) or external contamination during grinding. Though further studies can
 184 confirm the effect of soaking on arsenic uptake by wheat grain if any, we found a little stronger
 185 relationship (Figure 3b) between wheat flour and drinking water (Spearman's Rho= 0.3174,
 186 $P < 0.05$), often the same water used in the household for soaking compared to wheat grain
 187 (Spearman's Rho=0.2492, $P < 0.05$).

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189

190 **Figure 3:** a) Arsenic concentration in wheat flour compared to wheat grains; b) Arsenic
191 concentration in wheat flour and grain in relation to As in drinking water

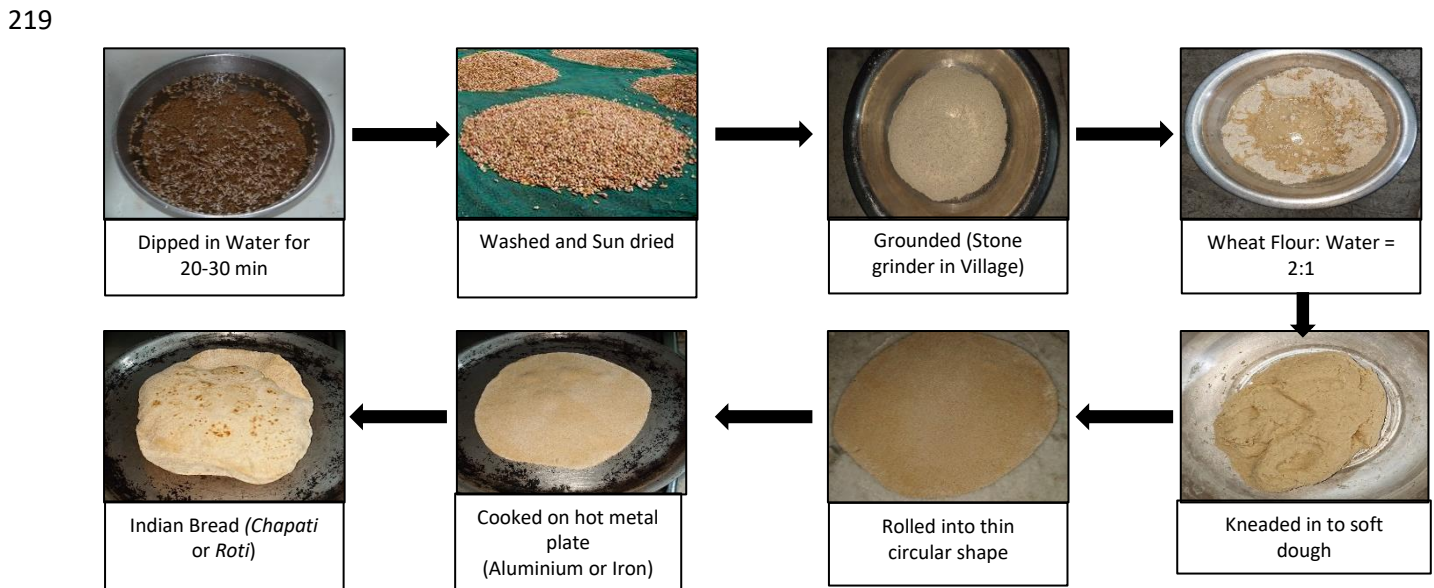
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193 ***2.2 Estimation of wheat intake from wheat-based food consumption***

194 Consumption of wheat was mainly from home-made bread (*chapati*), which is mostly cooked
195 in a traditional way in rural Bihar, as shown in Figure 4. The other wheat-based products
196 consumed by the surveyed population include roasted wheat-flour drink (*Sattu*), some bakery
197 products like biscuits and refined wheat products such as vermicelli, noodles etc. (Table 4).
198 Compared to a similar study in Pakistan (Rasheed et al., 2017), considering all potential dietary
199 sources of wheat, consumption (based on this study) in rural Bihar was found to be lower (Table
200 5).

201 In rural India, the percentage share of total daily energy intake from cereal grains is reported
202 to be 57.4% (NSS, 2014a), hence the diet is mainly dominated by cereals, which was found to
203 be true for rural Bihar with 67% of energy intake coming from cereals (NSS, 2014a). Next to
204 rice, wheat is the most important food-grain in India and is the staple food of millions of
205 Indians, particularly in the northern and north-western parts of the country. Per capita
206 consumption of wheat in Bihar (186 g/day) was reported to be higher than that for India (143
207 g/day) as shown in Table 5 (NSS, 2014b). While overall rural per capita consumption of wheat
208 in India has increased by 0.1 kg per person per month between 2004-05 and 2011-12, the
209 consumption via PDS (PDS stands for Public Distribution System, which means the
210 distribution of some essential commodities by the government at subsidised rates through
211 ration shops, fair price shops and control shops and more widely used in rural areas) has
212 increased considerably from 0.31 kg per person per month in 2004-05 to 0.74 kg per person
213 per month in 2011-12 (NSS, 2014b). Hence, there is an increase in wheat consumption in rural

214 India and furthermore in rural Bihar. Compared to other As affected areas of India, like West
 215 Bengal with subsistent rice-based diet, consumption of wheat-based food was found to be
 216 significant in rural Bihar. In fact, based on this study we found that 78% of the studied
 217 population consumed rice everyday while wheat-based home-made bread (*chapati*) was
 218 consumed everyday by all the participants (99.5%).



220 **Figure 4:** Household processing for preparation of Indian Bread (*Chapati* or *Roti*)

221

222 **Table 4:** Wheat-based food consumption in rural Bihar (mean \pm SD)

| | Total wheat intake | Indian bread (<i>Chapati</i>) | Bakery | Wheat flour drink (<i>Sattu</i>) | Refined Wheat Product |
|---|---------------------|------------------------------------|------------------|------------------------------------|-----------------------|
| Intake (g/day) | 223.78 \pm 107.59 | 215.07 \pm 109.29 | 2.48 \pm 10.00 | 1.59 \pm 9.90 | 3.61 \pm 13.40 |
| Percentage contribution to total wheat intake | | 94.90 \pm16.45 | 1.15 \pm 4.65 | 0.68 \pm 4.20 | 2.87 \pm 13.23 |

223

224 **Table 5:** Comparison of mean wheat intake in different studies

| Country/Area | Study Year | Age group/ Gender | Mean Wheat Intake (g/Day) | Reference |
|--------------|------------|----------------------|------------------------------|------------------------|
| Pakistan | 2017 | 3-6 Years | 149 | Rasheed et al. (2017) |
| | | 6-16 Years | 227 | |
| | | >16 Years Male | 426 | |
| | | >16 Years Female | 358 | |
| Bangladesh | 2004 | male | 179 | Watanabe et al. (2004) |
| | | female | 131 | |
| China | 2015 | children | 13 | Zeng et al. (2015) |
| | | adult | 44 | |
| Europe | 2013 | mean | 182 | FAO (2013) |
| India | 2011-12 | per capita | 143 | NSS (2014b) |
| Bihar | 2011-12 | per capita | 186 | |
| Bihar | 2017-19 | male | 272 | This Study |
| | | female | 165 | |

225

226 **3.3 Estimated increased cancer risk from wheat-based diet**

227 The input parameters defined as probability distributions are given in Table 6. In this studied
 228 population, 95% of wheat intake was from home-made bread (*chapati*) which is made from the
 229 wheat flour, hence total As concentration in the wheat flour was used over wheat grains in
 230 estimating the cancer risk from wheat intake (Table 6). Published literature on inorganic As
 231 content in wheat flour is sparse hence we have combined the results for both wheat grain and
 232 flour (Table 7) to determine the parameter for probabilistic risk estimation. Though we have
 233 used the bioaccessible fraction based on the study by Alhobiti and Beauchemin (2018), the
 234 bioaccessible fraction of As in wheat depends on the variety (Alhobiti and Beauchemin, 2018),

235 hence it should be further explored for Indian varieties. Thus, while the probabilistic risk
 236 assessment is limited by the accuracy and representativeness of the input data, it constitutes a
 237 framework to which improved data can readily be added to for improved risk estimation
 238 (Mondal and Polya, 2008).

239 Figure 5 presents the cumulative distribution of gender adjusted and separately for male and
 240 female, the excess lifetime cancer risk from wheat-based food intake. The point estimated mean
 241 risk of 1.08×10^{-4} was similar to probabilistic estimate of 1.23×10^{-4} , but the probabilistic
 242 approach to As risk assessment can take into account for the variabilities and parameter
 243 uncertainties (Mondal and Polya, 2008). Besides, the 5th percentile (9.36×10^{-6}) and 95th
 244 percentile (3.67×10^{-4}) could be relevant to determining the sub group of the studied population
 245 with low and high increased cancer risk from wheat-based food intake. Estimated risk in female
 246 (9.96×10^{-5}) was lower than in male (1.44×10^{-4}). Previously, for Pakistan, where wheat intake
 247 is comparatively higher (as shown in table 4), cumulative cancer risk was found to be 2 persons
 248 in a population of 10,000 in the highest exposed group (Rasheed et al., 2017) compared to our
 249 estimate of 1 in 10,000.

250 **Table 6:** The input parameters used in calculation of As attributable cancer risks

| Input variable | Point estimate | Probabilistic estimates | Data source |
|---|------------------|---|--|
| As in wheat flour ($\mu\text{g}/\text{kg}$) | 49.8 ± 74.08 | Pearsons (shape parameter =1.5, scale parameter=30.8) | This study |
| Inorganic As | 79% | Uniform ($a=29.48$, $b=106.75$) | Based on previous 6 studies (Table 7) |
| Bioaccessibility | 80% | Constant | Alhobiti and Beauchemin (2018) |

| | | | |
|---|--------------------------|--|---|
| Wheat intake (g/day) | male: 272 | Triang (Min=51.80, continuous mode=240, Max=556.81) | This study |
| | female: 165 | Normal (mean=170.58, SD=77.97) | |
| Body weight (kg) | male: 60 | Triang (Min=34.94, continuous mode=48.3, Max=96.26) | This study |
| | female: 52 | Weibull (scale parameter= 1.95, shape parameter= 23.45) | |
| Exposure duration (years) | equal to age if age < 40 | Constant | Mondal & Polya (2008) |
| | equal to 40 of age > 40 | | |
| Life expectancy (years) | male: 67.8 (Bihar) | Constant | https://niti.gov.in/content/life-expectancy |
| | female: 68.4 (Bihar) | | |
| cancer potency slope factor ((mg/kg)/d) ⁻¹ | 1.5 | Constant | USEPA |
| gender distribution (%) | male: 52.2% (Bihar) | Constant | Census of India (2011) |
| | female: 47.8% (Bihar) | Constant | |

251

252 **Table 7:** Inorganic As content in wheat

| Country | Year (published) | Sample | Sample Size | % inorganic As | Reference |
|---------|---------------------|--------------|----------------|-------------------|-----------------------|
| Italy | 2010 | Wheat Grains | 20 | 95.1 | Cubadda et al. (2010) |
| Italy | 2011 | Wheat Grains | 3 | 97 | D'Amato et al. (2011) |
| Italy | 2011 | Wheat flour | 3 | 95 | D'Amato et al. (2011) |

| | | | | | |
|-----------|------|---------------------------------------|----|-------|-----------------------|
| China | 2014 | Wheat Grains (Grown in low As) | 25 | 37.2 | Tong et al. (2014) |
| China | 2014 | Wheat Grains (Grown in High As) | 25 | 42.8 | Tong et al. (2014) |
| Argentina | 2016 | Wheat flour | 8 | 69.44 | Sigrist et al. (2016) |
| Pakistan | 2018 | Whole Wheat | 8 | 99.02 | Rasheed et al. (2018) |
| Belgium | 2018 | Wheat Grains | 9 | 79-94 | Ruttens et al. (2018) |

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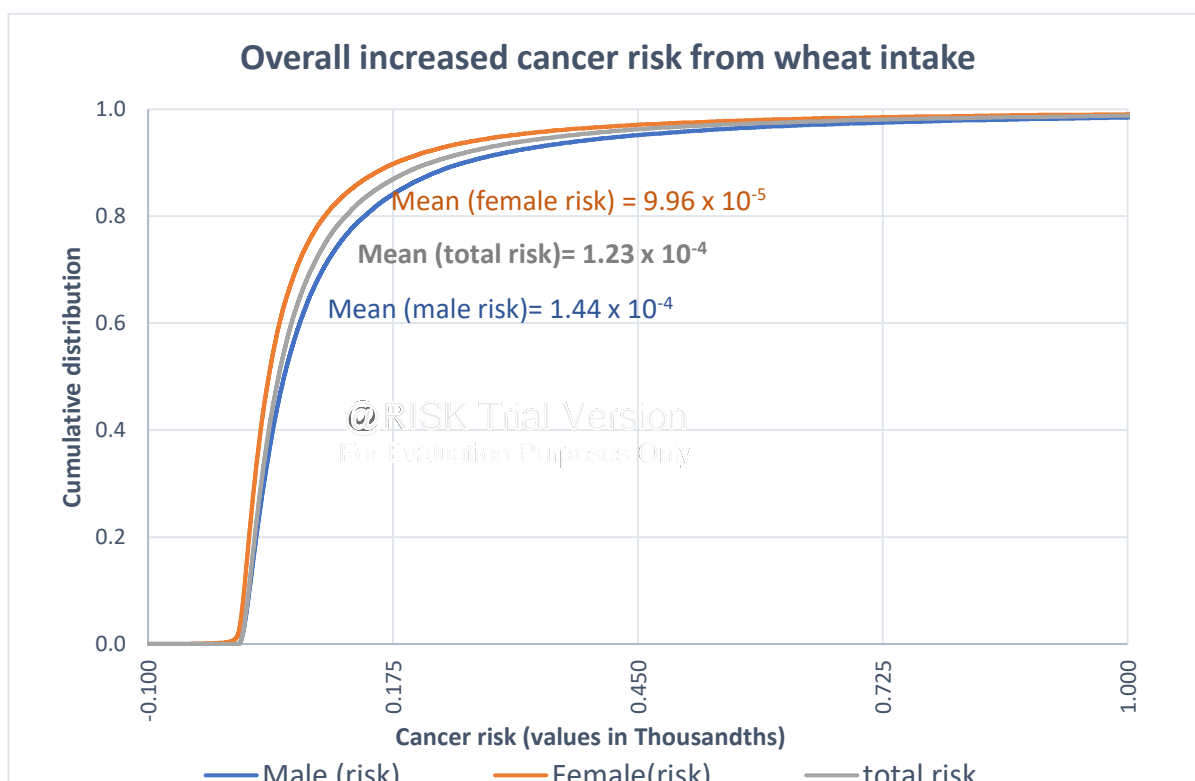
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263 **Figure 5:** Cumulative probability distributions of gender adjusted excess lifetime cancer risk

264 from wheat intake for the studied population.

265

266 **Conclusion**

267 Next to rice, wheat is the most important food-grain of India and As in wheat is an emerging

268 concern. To the best of our knowledge, this is the first study conducted using detailed dietary

269 assessment to determine As exposure from wheat-based food intake. The overall aim of this

270 study was to determine the increased cancer risk due to As from wheat-based diet in rural Bihar,
271 where As exposure is endemic.

272 Though total As estimated in wheat grains collected from the households in this study (43.64
273 $\mu\text{g}/\text{kg}$) was lower than reported As in wheat from other contaminated areas but it was higher
274 than a previous study from Bihar reporting a mean of 27 $\mu\text{g}/\text{kg}$ of total As. The estimated
275 excess lifetime cancer risk of 1.23×10^{-4} for the studied rural population of Bihar indicated risk
276 higher than the 10^{-4} - 10^{-6} range, typically used by the USEPA as a threshold to guide regulatory
277 values. Hence As exposure from wheat-based food intake is of concern in rural Bihar. In this
278 studied population, 95% of wheat was found to be consumed in the form of home-made bread
279 (*chapati*) prepared from the wheat flour which in turn was prepared from the wheat grains after
280 soaking them in water followed by drying in the open sun and grinding them to powder. In
281 contrast to previous studies, we did not find a significant decrease in total As in wheat flour
282 compared to wheat grains and rather a mean increase was observed, though not significant.
283 While the possible influence of the water used for soaking the wheat grain before grinding it
284 to flour and/or external contamination during grinding needs further investigation, our
285 observation confirms that modification of household-based wheat processing method may
286 reduce As exposure.

287 The As concentrations determined in the wheat samples collected from the households in this
288 study may not be representative of As in wheat cultivated in Bihar, since previous studies have
289 reported higher As concentrations in wheat cultivated in endemic areas. But this reinforces the
290 fact that due to the public distribution system in India, chances are there that the As
291 contaminated wheat is consumed in areas where there is no contamination in drinking water,
292 making it the most significant route of exposure, especially for sub-populations with similar
293 wheat-based diets as observed in this studied rural areas of Bihar. Moreover, since wheat-based

294 products are readily distributed, further studies should not just be restricted to As endemic areas
295 and should include As estimation in all wheat-based food products.

296 **Acknowledgement**

297 “Nature and nurture in arsenic induced toxicity of Bihar, India” (Nutri-SAM) is a UKIERI
298 project jointly funded by British Council in UK and Department of Science and Technology in
299 India (DST-UKIERI-2016-17-0064). SS is grateful to the School of Science, Engineering and
300 Environment, University of Salford for supporting his PhD. We acknowledge and thank Dr
301 Arun Kumar for his field support, and Kumar Bhaskar for his help with creating the map of the
302 study area. We are thankful to Professor David Polya, Yuru Bai, and Jingqi Zhang from
303 University of Manchester for their kind support with quality assurance of sample analyses. We
304 are grateful to all the participants of this study from villages of rural Bihar for providing us
305 with the samples. Instrument support from the GCER, The University of Newcastle is highly
306 appreciated.

307

308 **Competing interests**

309 The authors declare that they have no competing/conflicting interests.

310

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