1	Baseline distribution of petroleum hydrocarbon contamination in the marine environment
2	around the coastline of Qatar
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1 Abstract

2 Levels of organic contaminants (TPHs, PAHs) were simultaneously determined in both abiotic 3 (sediments, seawater) and biotic (Pinctada radiata oysters) samples at four sites along the coastline of Qatar (Arabian Gulf) in 2017-2018. TPHs and PAHs were more frequently detected 4 in oyster tissues than sediment and seawater samples collected from the same areas. While levels 5 6 of TPHs and PAHs in seawater and sediments were lower than previous local studies and worldwide studies, PAHs levels observed in pearl oyster tissue (25.9- 2240 µg/kg) were relatively 7 higher than in previous studies in Qatar. In general, eight PAHs compounds were detected in oyster 8 tissue, with benzo(a)pyrene displaying the highest concentration. The coast of Qatar could be 9 affected by seasonal patterns of pollutants, where TPHs and PAHs levels increased in winter 10 compared to summer. These results provide key information on the use of the pearl oyster as a 11 bioindicator species and Qatar's marine environment. 12

13 Keywords:

Total Petroleum Hydrocarbons, Polycyclic Aromatic Hydrocarbons, Bioindicator Species, *Pinctada radiata*, Bioaccumulation, Arabian Gulf

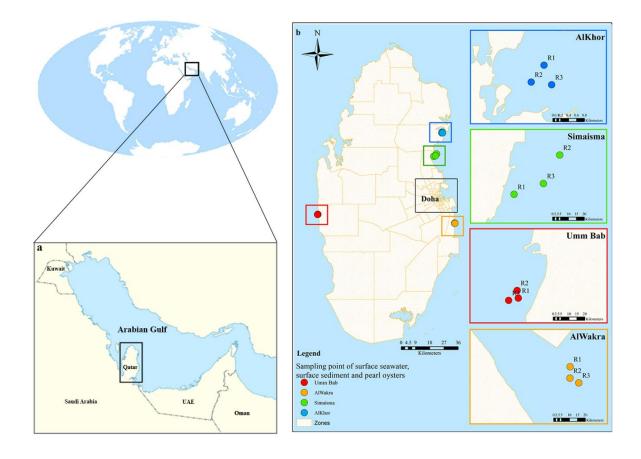
In recent decades, industry and population centers in the coastal areas of Qatar have expanded 1 rapidly (Shandas et al., 2017), impacting Qatar's marine environment. Impacts include wastewater 2 3 released to coastal waters and sediment resuspension caused by coastline development (Ministry 4 of Development Planning and Statistics, 2017). Some primary point sources of pollution have been 5 identified including the discharge of industrial and sewage effluent within a few kilometers of the 6 Qatari shoreline and wastes released from oil and desalination industries (Bejarano and Michel, 7 2010). These various activities have resulted in a significant influx of various chemical pollutants 8 in the marine environment of the Arabian/Persian Gulf, including some known to have a variety 9 of adverse (e.g., reproductive and mutagenic) impacts on marine organisms (Rushdi et al., 2017). Chemical contaminants have also been detected in seafood products from the Gulf, at levels 10 sometimes greater than international food safety limits (Freije, 2015). Qatar relies almost entirely 11 on seawater for the production of potable water (Ashfaq et al., 2019), and seafood is a key source 12 of protein in diet, hence, it is important to determine if these marine pollutants pose a threat to 13 national water and food security. Moreover, marine pollution is known to be a threat to ecosystems 14 and species diversity (Kurylenko & Izosimova, 2016). As such, protecting Qatar's marine 15 ecosystems, drinking water and food from chemical contamination has been identified as a core 16 17 component of the Environmental Development within the National Vision 2030 (MDPS, 2017).

Environmental risk assessments are now recognized to be best undertaken using an integrated approaches including both chemical and appropriate biological parameters in guardian species (EPA, 2019). As such, bivalve molluscs, including mussels and oysters, have been used extensively in marine coastal monitoring programs, from the early U.S. "Mussel Watch" program in the 1970s to more recent European research programs and initiatives in the Gulf in the context of the "*Regional Organization for the Protection of The Marine Environment*" (Tavares et al., 1988). Bivalves are ideal sentinel organisms, with high ecological importance and a wide

geographical distribution (Dumbauld et al., 2009). Furthermore, they can bio-accumulate 1 pollutants and exhibit a range of biological responses when stressed or exposed to environmental 2 3 contaminants (Melwani et al., 2013). In Qatar, the pearl oyster, Pinctada radiata is considered an 4 essential part of the nation's cultural heritage and one of the key economic foundations upon which 5 the country developed. Historically, the pearl oyster formed large beds on the western side of the 6 Arabian Gulf extending from Kuwait in the north to Oman in the South (Figure 1A). Due to its 7 biological characteristics, ecological status and sensitivity to anthropogenic stressors this symbolic 8 species of Qatar's maritime history provides an ideal sentinel species to monitor and assess the 9 marine environmental quality as has been studied by Liu et al. (2019); Fowler et al. (1993); Tolosa et al. (2005); and De Mora et al. (2010). Research work has been undertaken in several Gulf 10 countries to assess the contamination levels and bioavailability of organic contaminants in water, 11 sediment and biota tissues. For example, Freije (2015), Lyons et al. (2015), Beg et al. (2003), 12 Bach et al. (2003) and Saeed et al. (1996) reported the levels of Polycyclic Aromatic Hydrocarbons 13 (PAHs) in Kuwait's marine sediment. Tolosa et al. (2005) examined the total contamination level 14 of Total Petroleum Hydrocarbons (TPHs) and PAHs in biota and coastal sediments from Bahrain, 15 UAE, and Oman. TPH levels in marine sediment from Bahrain and oyster tissues from Oman have 16 been reported by De Mora et al. (2010), and there are studies on the bioavailability of TPHs in 17 sediments from Saudi Arabia and Kuwait (Fowler et al., 1993; Lyons et al., 2015). However, 18 although the Qatari marine environment has been impacted by marked industrial expansion and 19 20 anthropogenic pressure over the last several decades, except for specific studies on trace metals, few studies have investigated other pollutants such as TPH and PAH in Qatar's marine 21 environment. The aim of this study was to quantify the levels of organic contaminants (TPHs and 22 PAHs,) in abiotic matrices (seawater and sediments) and biotic tissues of *P. radiata* from different 23 locations in Qatar in different seasons; and to determine the relationships (correlations) between 24

- 1 the organic contaminants among different abiotic and biotic matrices. The applicability of *P*.
- 2 *radiata* as an early warning pollution indicator for organic contaminants was also explored.

Four coastal sampling sites were selected (Figure 1B) to represent different anthropologically 3 impacted areas. Al-Khor was selected as a relatively pristine inshore site with water depths of 0.5-4 1.0 m (Leitão et al., 2017). The area is relatively free from industries and urbanization, apart from 5 6 occasional recreational activity. Al-Wakra, situated in eastern Qatar, 15 km south of the capital Doha, is an inshore location with a water depth of 0.5-1 m with sandy bottom sediment which is 7 8 very busy with recreational boat traffic and fishing fleet movement. There is a boat yard where minor repairs are carried out on seagoing vessels, fueling activities are also found within the area 9 where fishing and recreational vessels are refueled. Umm Bab is located in western Qatar, only 25 10 11 km from the industrial city of Dukhan, and hosts a seawater desalination plant near the coast and a cement company. Simaisma is located on the east coast of Oatar with a water depth of 1-5 m. 12 The area situated within a small bay approximately midway between Lusail and Al Khor. Samples 13 were collected from the four different sites in March 2017, December 2017, May 2018 and 14 November 2018. 15



2 **Figure 1:** Location of Qatar in the Arabian Gulf (a) and the sampling sites in the Qatar coastal zone (b).

3

Seawater, sediment and oyster batches were collected in triplicates from each site (Al Khor, 4 Simaisma, Umm Bab, Al Wakra – See Figure 1) over four rounds (March 2017, December 2017, 5 6 May 2018 and November 2018), resulting in a total number of 12 samples for each matrix from 7 each site. Seawater samples were collected from the sea surface using pumps and seawater 8 temperature, salinity and pH were measured in the field using an EXO3 Multiparameter Sonde (YSI). Composite sediment samples were collected using a Van-Veen grab (0.05 m2). TOC 9 10 content in seawater, sediment mean organic content and particle size were measured in the 11 laboratory (See Supplemental Methods). P. radiata oysters were collected (n=3 composites of >12 oysters per site for each sampling round) by scuba divers at the targeted sites, where water depth 12 did not allow the direct collection of samples. The size (long axis) was recorded for each organism 13

1 and ranged from 2 to 7 cm. Specimens were transported to the laboratory in seawater and were 2 transferred to a fridge (4°C) prior to processing. Oysters were subsequently opened, rinsed with 3 ultrapure water and dissected over ice. In total, 576 oysters were analyzed (48 batches). The levels 4 of TPHs (C₈ to C₃₂) and 16 PAHs compounds, listed as priority pollutants by the US EPA, were 5 measured following US EPA Methods 3510C for seawater and Method 8015B for surface marine 6 sediment and oyster tissue. Gas Chromatography / Flame Ionization Detector (GC/FID) and Gas 7 Chromatography/ Mass Spectrometry (GC/MS) were used for the analysis of TPH and individual 8 PAHs (Supplemental Table S1-S2). Quality assurance included duplicate samples and analyses 9 of standard reference materials ("CRM361 TPH - Sea Sediment" from Fluka Analytical and "SRM 1941b Organics in Marine Sediment" from NIST, for TPHs and PAHs respectively). Recoveries 10 ranged 80-120% (Supplemental Table S3). The statistical analyses were performed using Excel 11 and SPSS software. The Normality test was applied to all data to check the distribution of data. 12 For normal distributions, One-way ANOVA was performed followed by a Tukey HSD (Tukey's 13 Honestly-Significant Difference) post-hoc test to determine which pairs of the factor were 14 significantly different from each other. Whereas for data that was not a normal distribution, the 15 nonparametric test (Kruskal-Wallis test) was performed. Post-hoc pairwise comparisons were only 16 applied where a significant difference was noted. The LOD divided by the square root of 2 was 17 used for samples below the limit of detection. 18

The physicochemical parameters and contaminant levels in seawater are presented in Table 1 for the four sites and four sampling rounds (PAHs values are not presented in this table because they were negligible). The surface seawater samples from the four different locations showed negligible levels of PAHs, while TPHs ranged from 51.9 to 89.7 μ g/L. Low concentration of PAHs observed in seawater samples collected in this study may be due to relatively high hydrophobicity of PAHs (Karthikeyan et al., 2001). In general, TPHs concentrations in seawater in the spring / summer

1	month (March 2017 and May 2018) were higher than the winter months (December 2017 and
2	November 2018) (see supplemental Table S5). The observed TPHs concentrations in seawater
3	(mean 0.160 to 486 μ g/L) were lower than those found in a previous study in Qatar (30 to 1700
4	μ g/L) (Leitão et al., 2017). While detectable levels of PAHs (0.06 to 10.76 μ g/L) was observed
5	earlier by Leitão et al. (2017), negligible levels of PAHs was noted in this study. A review of the
6	literature from the Gulf region and worldwide environmental studies showed that our result are in
7	line and sometimes lower than most findings (Saeed et al., 2011;El Nemr and Abd-Allah, 2003,
8	Xiang et al., 2018 and Nizzetto et al., 2008; Valavanidis et al. 2008; Ţigănuș et al., 2016 Kim et
9	al. 2013).

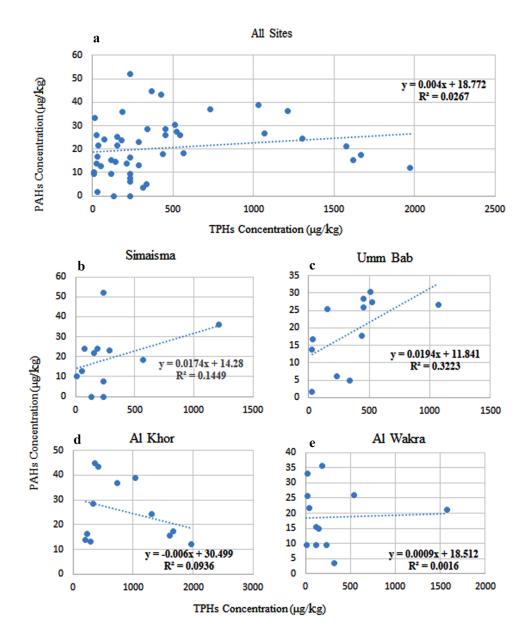
Site **TPHs** Temperature Salinity TOC pН (mg/L) $(\mu g/L)$ (psu) (°C) Simaisma $Mean \pm SD$ $24.4\pm\!\!3.10$ 8.17 ± 0.05 43.0 ± 0.7 27.8 ± 0.7 51.9 ± 84.1 20.4 8.08 2.70 42.1 26.7 Min. 8.24 250 29.2 44.6 28.5 Max. 52.6 ± 68.9 24.8 ± 3.80 8.01 ±0.16 56.1 ±2.2 29.5 ± 0.7 Umm Bab $Mean \pm SD$ 19.0 7.81 52.6 28.7 1.90 Min. 29.2 8.24 58.2 30.9 190 Max. 25.3 ± 4.6 8.20 ± 0.12 43.3 ± 0.5 28.3 ± 0.9 89.7 ± 139 Al Khor Mean± SD 18.0 7.99 42.3 26.6 1.00 Min. 30.4 8.33 44.3 29.9 486 Max. 27.0 ± 2.00 8.11±0.08 41.4 ± 1.8 28.4 ± 1.5 65.1 ± 64.8 Al Wakra Mean 23.8 7.92 38.5 25.6 0.160 Min. 29.2 8.18 170 43.5 30.3 Max.

Table 1. Physicochemical parameters and contaminant levels in seawater samples (n=12 for each site)

12	Significant differences were observed between the percentage of sand, silt and clay in the surface
13	sediments and sites ($p < 0.05$; Supplemental Table S4). Surface sediments were identified as sand
14	in Umm Bab and Al Wakra, and as silty sand in Simaisma and Al Khor. Sandy sediment generally
15	has less retention with regards to TPH and PAH. The mean OC content in surface marine sediments
16	from the four sites in Qatar ranged from 0.079 to 0.219 % (Supplemental Table S4). Relatively

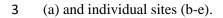
higher sediment OC was observed in Umm Bab and Al Khor in comparison to Simaisma and Al 1 Wakra, though this difference was only significant for Umm Bab (p=0.002). The mean 2 3 concentration of TPHs ranged between 75.0 to 1750 µg/kg in surface sediments (Figure 2, 4 Supplemental Table S5). Significantly higher TPHs concentrations were observed in Al Khor 5 (p=0.01) compared to Umm Bab, Simaisma and Al Wakra which was not expected as all area have 6 similarly industries. The average values of the Σ_{16} PAHs concentrations in surface marine 7 sediments from various stations were between 4.25 to 36.7 µg/kg (Supplemental Table S6). The 8 highest concentration was found in Al Khor, followed by Simaisma, then Al Wakra with lowest 9 levels at Umm Bab. From all the 16 tested compounds, only six were detected in the marine sediment of which pyrene was the most prominent within all sites (mean 10.1 µg/kg), 10 benzo(a)pyrene (B[a]P; mean 4.44µg/kg) was detected mostly in winter while indeno(1,2,3-11 cd)pyrene (mean 2.23µg/kg) was detected within the summer month in all sites. There was a weak 12 positive correlation between the levels of TPHs and levels of PAHs in marine sediment samples 13 collected from Qatar's marine environment (Figure 2-a). While a stronger positive correlation was 14 observed at Umm Bab (Figure 2-c) followed by Simaisma (Figure 2-a), no relationship could be 15 observed at Al Wakra (Figure 2-e) and surprisingly a negative, though weak correlation, was noted 16 17 for Al Khor (Figure 2-d).

In this study, TPHs measurement ranged from 75.02 to 1751.82 μ g/kg whereas the range in previous local studies of marine sediment from Qatar spanned 17.3- 79.3 μ g/kg (Rushdi et al., 2017); 160- 1700 μ g/kg (Leitao et al., 2017); and 2200- 84000 μ g/kg (Tolosa et al., 2005). The high levels reported by Tolosa et al. (2005) where from Qatari sites impacted by the oil terminal and refinery plant (Umm Said, Dukhan, Ras Lafan). Moreover, the concentrations of total PAHs ranged from 4.25 - 36.7 μ g/kg and were in the same range (Leitao et al., 2017; Soliman et al. 2019). Σ PAHs reported in Qatar's marine sediments were low compared to data from the wider Gulf (7200- 80000 μg/kg, Readman et al., 1996; 1333.6 μg/kg, Beg et al., 2003; 12.9-190 μg/kg, Lyons
et al., 2015). Moreover, concentrations of ΣPAHs from Qatar's marine sediments were much lower
than those reported in other international studies conducted in China, Russia and the UK (13.59166.5 μg/kg, Zhang et al., 2016; 61.2-368 μg/kg, Readman et al., 4171-79648 μg/kg, 2002 Brettell,
2013). This may be related to the high proportion of sand in the sediments in the current study.





2 Figure 2: Correlation between TPHs and PAHs "µg/kg dry weight (dw)" in surface sediments for all sites



1 Mean TPH content in oysters ranged from 633 to 6670 μ g/kg with the highest concentrations 2 measured in Simaisma. There was a significant seasonal difference in the TPHs in oyster tissue 3 samples at Simaisma and Al Wakra. The mean concentrations of PAHs in oyster tissue samples 4 showed an extreme variation from 25.9 to 2240 µg/kg. Significantly high levels of PAHs were 5 measured in Al Wakra (p <0.05). In general, ten PAHs compounds were detected in oyster tissues 6 (Supplemental Table S7). Benz(a)pyrene was the highest PAH concentration detected (1510 7 µg/kg) observed in Al Wakra. BaP is a known carcinogenic compound and has recently been flagged by the International Agency for Research on Cancer (IARC). Another carcinogenic, 8 9 dibenz[a,h]anthracene, was detected in Al Wakra samples collected in winter seasons (both 2017 and 2018). In general, fluctuations in the patterns of PAHs in oyster tissues suggest that the coast 10 of Qatar could be affected by seasonal patterns of pollutants. Notably, PAHs levels increased in 11 winter compared to summer. 12

Considering all the sites, TPHs and PAHs concentrations in oyster tissues showed a weak positive 13 14 association (Figure 3). However, at Simaisma and Al Khor there was a strong, positive, linear correlation between the two organic contaminants ($R^2=0.815$ and $R^2=0.885$ respectively). TPHs 15 concentrations reported in *P. radiata* tissues in this study (633- 6670 µg/kg) were in the range of 16 previous study from Qatar (10 to 7410 µg/kg) (Leitão et al., 2017). TPHs concentrations in other 17 Gulf countries were significantly higher than levels found in this study and previous studies in 18 Qatar (De Mora et al. 2010; Tolosa et al. 2005). Levels reported ranged from (22900 to 69700 19 20 μ g/kg) in Bahrain, (20300 to 33900 μ g/kg) in UAE, and (37600 to 632000 μ g/kg) in Oman as described by de Mora et al. (2010). Whereas Tolosa et al. (2005) reported high levels of TPHs in 21 22 oyster tissues in the same Gulf countries, concentration ranges were (38500 to 98000 µg/kg, 25600 to 237000 µg/kg, and 23200-130000 µg/kg) in Bahrain, UAE and Oman, respectively. 23 Furthermore, high TPHs range were observed in different areas around the world. Vaezzadeh et 24

al. (2017) studied TPHs in oyster tissues from Malaysia and reported a level from 56700 to 263000
µg/kg, while Inam et al., 2012 reported levels in the range of 6370- 8440 µg/kg in oyster tissues
from Nigeria. Overall, the organic contaminants (TPHs and PAHs) were more readily detected in
oyster tissue samples than marine sediment and seawater samples collected from the same areas
(Figure 4, Table 1).

Levels of PAHs observed in this study in P. radiata oyster tissue samples (25.9-2240 µg/kg) were 6 higher than the previous observation by Leitão et al. (2017) who reported 0.150-23.1 µg/kg in P. 7 radiata tissues from Qatar's marine environment. The levels of PAHs reported in this study were 8 close to levels reported in other Gulf countries. De Mora et al. (2010) reported PAHs concentration 9 in oyster tissues between 80.0- 335 µg/kg from Bahrain, 19.0- 161 µg/kg from UAE, and 103-10 13700 µg/kg from Oman. These levels were slightly higher than a study reported five years earlier 11 in same areas, where the PAHs ranged from 58.3-105, 36.6-251, and 17.0-173 µg/kg in Bahrain, 12 UAE and Oman respectively (Tolosa et al., 2005). Concentrations of PAHs noted in P. radiata 13 14 oyster tissue samples collected from Qatar marine environment appear to exceed some other countries at least based on the highest values reported (Hong et al., 2016; Barhoumi el al., 2016; 15 Acosta et al., 2015; Nakata et al., 2014; Ramdine et al. 2012, Oros and Ross, 2005). 16

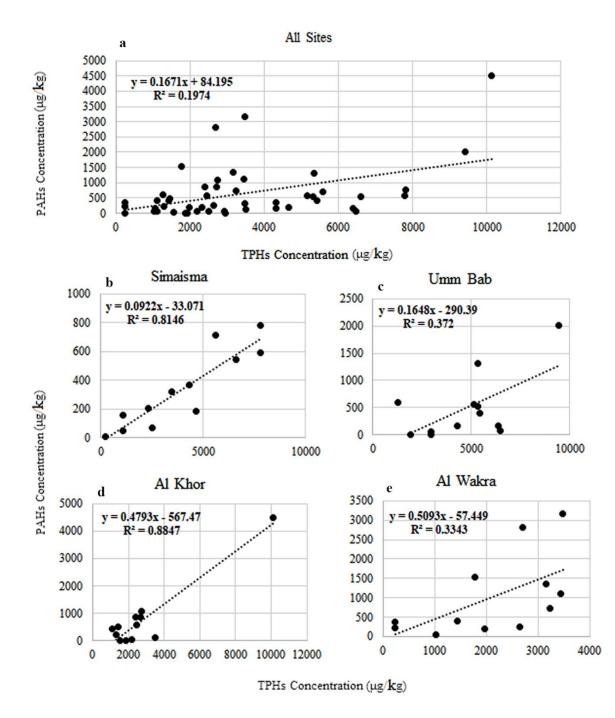


Figure 3: Correlation between TPHs and PAHs "µg/kg dry weight (dw)" in oyster tissues (all seasons) for
all sites (a) and individual sites (b-e).

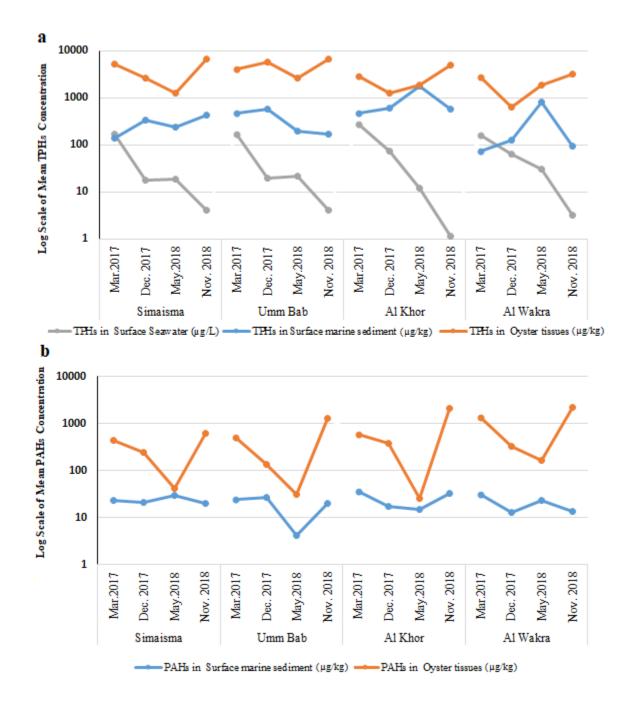


Figure 4: Seasonal variations of TPHs (a) and PAHs (b) in seawater (grey), sediments (blue) and oysters
 (orange). (Note: PAHs were not detected in seawater)

Based on our findings, while levels of TPHs and PAHs in seawater and sediment samples were 1 2 lower than previous local studies and examples worldwide, the PAHs levels observed in pearl 3 oyster tissue samples (25.9-2240µg/kg) were higher than previous studies in Qatar and even higher 4 than levels recorded in different areas around the world based in our selected studies. The Al Khor 5 site showed a higher range of TPHs concentration in seawater samples (1.20-272 μ g/L) than in a 6 previous local study (40.0-50.0 µg/L), and higher levels of TPHs and PAHs in surface sediment 7 than other three sites (480- 1750 μ g/kg and 15.0- 36.7 μ g/kg respectively). On the other hand, the 8 highest PAHs ranges in oyster samples were observed from Al Wakra and Al Khor (167-2240 9 $\mu g/kg$, and 25.9- 2080 $\mu g/kg$ respectively) with higher levels reported in winter season than summer. Therefore, although the sediments did not seem to have been affected by organic 10 contaminates, high levels of PAHs recorded in oysters collected from Al Wakra and Al Khor, and 11 in seawater collected from Al Khor might reflect residual hydrocarbons in these east coast sites of 12 Qatar. Al Wakra reported the highest level of PAHs, which may suggest chronic contamination 13 due to the boat traffic, fishing or fueling activities around this location. The present study provides 14 useful background information on the use of the pearl oyster as a bioindicator species and for 15 further investigations to understand the presence of organic contaminants in Qatar's marine 16 17 environment.

1 Acknowledgments

We would like to thank the Environmental Science Center (ESC) at Qatar University for the use
of their labs. This publication was made possible by National Priorities Research Program (NPRP)
award [NPRP9-394-1-090 "The Pearl Oyster: from national icon to guardian of Qatar's marine
environment"] from the Qatar National Research Fund (a member of the Qatar Foundation). The
statements made herein are solely the responsibility of the authors.

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